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Substantial hysteresis in emergent temperature sensitivity of global wetland CH₄ emissions

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Wetland methane (CH₄) emissions (F_{CH_4}) are important in global carbon budgets and climate change assessments. Currently, F_{CH_4} projections rely on prescribed static temperature sensitivity that varies among biogeochemical models. Meta-analyses have proposed a consistent F_{CH_4} temperature dependence across spatial scales for use in models; however, site-level studies demonstrate that F_{CH_4} are often controlled by factors beyond temperature. Here, we evaluate the relationship between F_{CH_4} and temperature using observations from the FLUXNET-CH₄ database. Measurements collected across the globe show substantial seasonal hysteresis between F_{CH_4} and temperature, suggesting larger F_{CH_4} sensitivity to temperature later in the frost-free season (about 77% of site-years). Results derived from a machine-learning model and several regression models highlight the importance of representing the large spatial and temporal variability within site-years and ecosystem types. Mechanistic advancements in biogeochemical model parameterization and detailed measurements in factors modulating CH₄ production are thus needed to improve global CH₄ budget assessments.

Methane (CH_4) is the second most important climate forcing trace gas influenced by anthropogenic activities after carbon dioxide (CO_2)^{1–3}. Wetlands are the largest and most uncertain natural CH_4 source, contributing 19–33% of current global terrestrial CH_4 emissions (F_{CH_4})^{4–6}. Top-down estimates from atmospheric inversion models and bottom-up estimates from in situ measurements both indicate gradual increases in natural wetland F_{CH_4} from 2000 (147–180 Tg CH_4 yr^{−1}; bottom-up vs. top-down) to 2017 (145–194 Tg CH_4 yr^{−1}), although F_{CH_4} estimates from both approaches vary widely^{4,6}. In addition, atmospheric CH_4 concentrations have rapidly increased since 2007 ($+6.9 \pm 2.7$ ppb CH_4 yr^{−1} for 2007–2015 vs. $+0.5 \pm 3.1$ ppb CH_4 yr^{−1} for 2000–2006), with increases arising from both biogenic (primarily agriculture and waste sectors) and fossil fuel-related sources^{7,8}. Observed atmospheric CH_4 concentrations have risen consistently with RCP8.5 (Representative Concentration Pathway of 8.5 W m^{−2})⁹ projections since 2007, and are growing relatively faster than observed increases in CO_2 concentrations during the same period⁸.

Wetland F_{CH_4} estimates are poorly constrained due to high temporal and spatial variability^{10,11}, compounded by insufficient measurements of fluxes (e.g., latitudinal data bias) and predictor variables (e.g., soil temperature and moisture), knowledge gaps in CH_4 biogeochemistry¹², and incomplete process representation in biogeochemical models^{4,5,13–15}. Several factors have been suggested to regulate wetland F_{CH_4} through effects on methanogenesis (i.e., production), methanotrophy (i.e., oxidation), and CH_4 transport, including gross primary productivity (GPP)¹⁶, water table depth (WTD)¹⁷, vegetation composition^{18,19}, redox conditions²⁰, substrate quality and availability^{21,22}, pore water CH_4 solubility²³, microbial community dynamics and activity²⁴, and temperature²⁵. At ecosystem scale, some in situ observations indicate that F_{CH_4} are mainly controlled by 20–35 cm depth soil temperatures and are not sensitive to WTD variations as long as anoxic conditions exist^{26–28}. Although F_{CH_4} appears to be positively correlated with temperature and CH_4 production^{24–27}, how to parameterize CH_4 production, oxidation, and emission rates in models remain key uncertainties. Reducing the uncertainties is required to improve global CH_4 budget assessments and increase confidence in future climate projections, as the temperature sensitivity of CH_4 biogeochemistry is parameterized differently among CH_4 models^{13,14,29}. A recent meta-analysis reported that CH_4 production temperature sensitivities derived from laboratory cultures are consistent with those of F_{CH_4} inferred from ecosystem-scale measurements and could therefore be used as an empirical basis for F_{CH_4} temperature sensitivity in models³⁰.

However, site-specific emergent F_{CH_4} temperature dependencies inferred from different measurement periods show substantial intra-seasonal variability over the course of the year^{31–33}, highlighting effects from other environmental drivers. For example, intra-seasonal variability may stem from hysteretic (i.e., temporally offset) microbial and abiotic interactions³⁴: higher substrate availability increases methanogen biomass and CH_4 production and emission later in the frost-free season³³. Similarly, higher F_{CH_4} for a given GPP later in the frost-free season has been reported, which may be caused by the time required to convert GPP to methanogenesis substrates²⁶. Further, changes in WTD can regulate the emergent F_{CH_4} temperature sensitivity through controls on soil redox potential^{31,35–37}, especially when the WTD is below the site-specific rooting depth and critical zone of CH_4 production^{17,38,39}.

Here, we evaluated observationally based emergent relationships among F_{CH_4} , GPP, WTD, and air (T_{air}) and soil (T_{soil}) temperatures using the global FLUXNET- CH_4 database⁴⁰. We analyzed data recorded in eight ecosystem types: bog, fen, marsh,

peat plateau, rice paddy, salt marsh, swamp, and wet tundra that spans 207 site-years across 48 wetland and rice paddy sites (Supplemental Fig. 1 and Supplemental Table 1). The FLUXNET- CH_4 database provides half-hourly ecosystem-scale eddy covariance measurements of F_{CH_4} and other fluxes (e.g., CO_2 , water vapor, and energy) measured at 83 sites across the globe⁴⁰ (including uplands, wetlands, and rice paddy sites). Apparent F_{CH_4} hysteresis has been observed in response to WTD^{17,31}, GPP²⁶, T_{air} ³³, and T_{soil} ^{27,32,33} at individual sites, but has not been synthesized across ecosystem types over distinct climate zones. Here, we analyzed intra-seasonal changes in emergent dependencies of F_{CH_4} on these potential controls at each site-year. We focused on relationships of F_{CH_4} with T_{air} because T_{air} is directly relevant to climate policy and better characterized in climate models⁴¹. In addition, the amount of T_{air} data (207 site-years) in the FLUXNET- CH_4 database is about twice than that of T_{soil} measured at the shallowest (0–18.3 cm; 112 site-years) and deepest (32–50 cm; 97 site-years) site-specific soil depths. We show that consistent intra-seasonal changes in emergent dependencies of F_{CH_4} were derived with T_{air} and T_{soil} measurements at the sites where both measurements were available.

We quantified emergent F_{CH_4} – T_{air} dependencies using a quadratic relationship (Methods; Eq. 1) fit to daily measurements reported during the frost-free season (defined by $T_{\text{air}} > 0^\circ\text{C}$, Methods). This quadratic functional form was chosen because it is consistent with MacroMolecular Rate Theory³³ analyses of the temperature sensitivity of CH_4 production and oxidation³⁴ and produced reliable estimates of F_{CH_4} for our study sites (Supplemental Fig. 2). For each frost-free season, seasonal F_{CH_4} hysteresis was quantified as changes in emergent F_{CH_4} – T_{air} dependencies inferred from earlier and later periods separated by the maximum seasonal T_{air} . We did not consider F_{CH_4} outside the frost-free season, although they can be important in some high-latitude wetlands^{32,42}. We used two metrics to quantify intra-seasonal changes in emergent F_{CH_4} – T_{air} dependence: (1) Normalized area of seasonal F_{CH_4} hysteresis (H_A ; i.e., the area enclosed by emergent earlier and later period F_{CH_4} – T_{air} relationships (Fig. 1d) normalized by maximum seasonal F_{CH_4} and T_{air} ; Methods); and (2) Mean seasonal F_{CH_4} hysteresis (H_μ ; i.e., the difference between mean daily F_{CH_4} inferred from measurements taken between later and earlier periods of the frost-free season). These two metrics are conceptually similar to those used to quantify temperature hysteresis in soil respiration⁴³ and soil CO_2 concentrations⁴⁴. Positive and negative H_A and H_μ values represent higher (e.g., Fig. 1d) and lower (e.g., Supplemental Fig. 3d) F_{CH_4} later (i.e., after reaching maximum seasonal T_{air}) in the frost-free season, respectively.

Results and discussion

A case study of positive seasonal CH_4 emission hysteresis. As an example of seasonal hysteresis, we examined daily estimates obtained from measurements taken at the Bibai Mire in Northern Japan (JP-BBY) where F_{CH_4} is insensitive to the relatively shallow WTD from 2015 to 2017²⁷ (Fig. 1b, c). Although the seasonality shown in F_{CH_4} appears to follow T_{air} (Fig. 1a, b), a time-dependent F_{CH_4} – T_{air} relationship varies from earlier to later parts of the frost-free season (Fig. 1d–f). Specifically, plotting daily F_{CH_4} as a function of T_{air} results in a counterclockwise loop from beginning to end of the frost-free season. Similar hysteretic patterns were found using T_{soil} (Supplemental Fig. 4) and gap-filled CH_4 emissions⁴⁵ (Supplemental Fig. 5), indicating that the hysteresis is not caused by time lags between T_{soil} and T_{air} resulting from heat transfer into the soil⁴⁶, and is not driven by biases

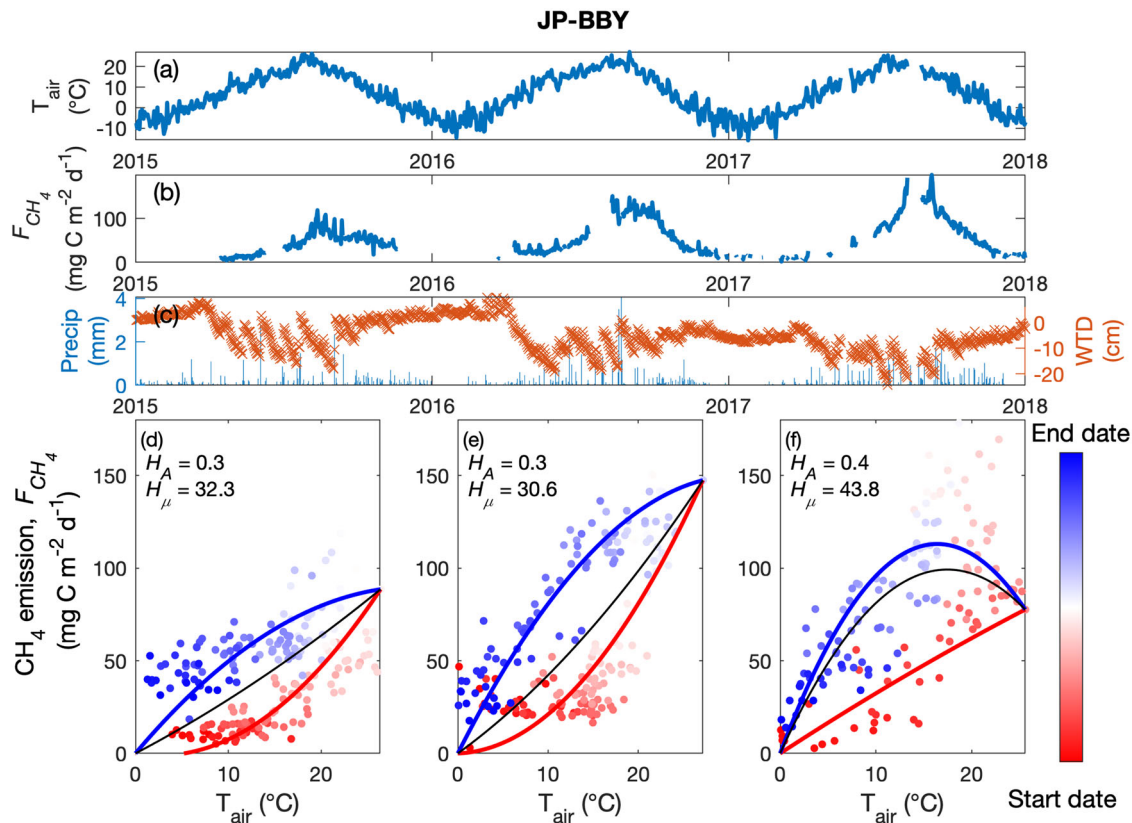


Fig. 1 Daily mean CH_4 emissions have hysteretic responses to air temperature. The quality-controlled daily air temperature (a), CH_4 emissions (b), precipitation (c, left axis), and water table depth (c, right axis) measured at the Bibai Mire in Japan (JP-BBY) from 2015 to 2017. CH_4 emission-air temperature dependencies (lines) derived from daily estimates (dots) recorded at JP-BBY for 2015 (d), 2016 (e), and 2017 (f). The results inferred from earlier and later parts of the frost-free season, and full frost-free season are colored in red, blue, and black, respectively. Start and end dates represent the beginning and ending of the frost-free season, respectively. Values of H_A and H_μ denote the normalized area of seasonal CH_4 emission hysteresis (normalized area enclosed by the blue and red lines) and the mean seasonal CH_4 emission hysteresis calculated in each site-year, respectively.

caused by missing data. These hysteretic patterns suggest that F_{CH_4} should not be represented as a single static function of T_{air} .

Seasonal CH_4 emission hysteresis among site-years. Overall, we detect positive seasonal F_{CH_4} hysteresis in most site-years recorded in the FLUXNET- CH_4 database, both in terms of H_A and H_μ (75–77% of site-years; Fig. 2). Consistent hysteresis patterns and magnitudes were found with monthly F_{CH_4} and T_{air} estimates (72–74%, Supplemental Fig. 6), indicating the observed seasonal F_{CH_4} hysteresis is not sensitive to temporal resolution. The non-zero H_A and H_μ values demonstrate intra-seasonal changes in emergent F_{CH_4} - T_{air} dependencies among wetland and rice paddy sites across the globe, and their negatively skewed distribution indicates that the hysteretic responses are not likely to be random. Ignoring seasonal F_{CH_4} hysteresis leads to overestimated ($28 \pm 46\%$) and underestimated ($-9 \pm 35\%$) F_{CH_4} predictions earlier and later in the frost-free season across wetland and rice paddy sites, and such prediction bias is overlooked by using seasonally invariant T_{air} dependence models ($-4 \pm 7\%$, Supplemental Fig. 7). For example, F_{CH_4} predictions made by a seasonally invariant emergent F_{CH_4} - T_{air} dependence at JP-BBY (i.e., black lines in Fig. 1d–f) are generally biased high and low in the earlier and later parts of the frost-free season, respectively.

To examine how potential controls are related to the observed seasonal F_{CH_4} hysteresis, we analyzed the distribution pattern of H_A under different site classifications and microclimatic conditions. The majority of site-years show positive seasonal F_{CH_4}

hysteresis when H_A values are categorized into (1) different ranges of mean T_{air} measured in the frost-free season (Supplemental Figs. 8), (2) different wetness conditions indicated by higher and lower mean WTD later in the frost-free season (Supplemental Fig. 9), and (3) different ecosystem types (Supplemental Fig. 10). Intra-seasonal changes in emergent GPP- T_{air} dependencies show about equal site-year proportions of positive and negative H_A values (48% and 52%, respectively; Supplemental Fig. 11a), suggesting that GPP does not directly contribute to the observed seasonal F_{CH_4} hysteresis. Further, predominantly positive seasonal F_{CH_4} hysteresis is detected using T_{soil} measured at the shallowest (Supplemental Fig. 12) and deepest (Supplemental Fig. 13) site-specific soil layers, indicating substantial intra-seasonal variability in the F_{CH_4} - T_{soil} relationship. Overall, the wetland and rice paddy observations in the current FLUXNET- CH_4 database suggest that F_{CH_4} are generally higher later (i.e., after reaching maximum seasonal T_{air} or T_{soil}) in the frost-free season at a given T_{air} and T_{soil} . These hysteretic responses emerged across climate zones with various GPP and frost-free season lengths, and were not directly attributable to intra-seasonal changes in T_{air} and T_{soil} (Supplemental Fig. 14).

Divergent temperature responses among sites and years. In terms of the magnitude of seasonal F_{CH_4} hysteresis, intra-seasonal changes in emergent F_{CH_4} - T_{air} dependence vary substantially among site-years within each ecosystem type (Fig. 3), despite being predominantly positive (Fig. 2). For each ecosystem type,

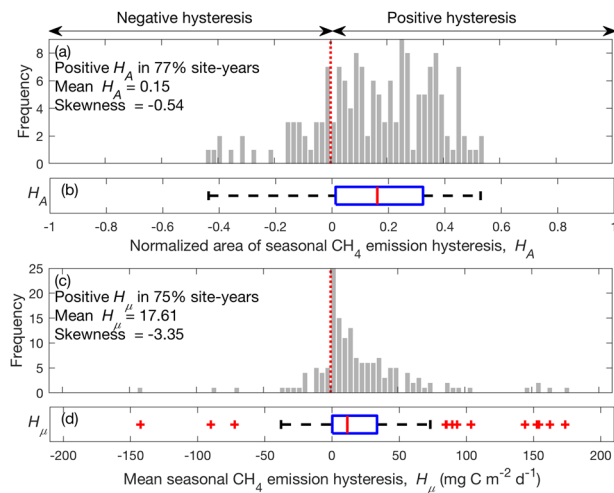


Fig. 2 Predominantly positive seasonal CH₄ emission hysteresis inferred from ecosystem-scale measurements across the globe, i.e., CH₄ emissions are generally higher later in the frost-free season at the same temperature. The distribution of normalized area of seasonal CH₄ emission hysteresis (H_A ; **a**, **b**) and mean seasonal CH₄ emission hysteresis (H_μ ; **c**, **d**) to air temperature among site-years derived from the FLUXNET-CH₄ database. Positive seasonal CH₄ emission hysteresis indicates higher CH₄ emissions later in the frost-free season at the same temperature (e.g., Fig. 1d–f). Red dashed lines represent no hysteresis. The corresponding boxplot of site-year specific H_A (**b**) and H_μ (**d**) derived from the FLUXNET-CH₄ database. The red central mark, and the bottom and top edges of the blue box indicate the median, and the 25th and 75th percentiles, respectively. The black whiskers extend to the most extreme data points not considered outliers denoted in red plus symbol.

the large inter-annual (i.e., different years within the same site) and inter-site (i.e., different site-years within the same ecosystem type) variability highlights the challenge of quantifying a universal and robust emergent F_{CH_4} – T_{air} dependence across wetland and rice paddy sites. For example, using the Boltzmann–Arrhenius function (Methods) to represent the emergent F_{CH_4} – T_{air} dependence of an ecosystem type cannot accurately reflect the site- and time-specific emergent relationships between F_{CH_4} and T_{air} (Fig. 3). A single static function of T_{air} thus cannot provide accurate estimates of F_{CH_4} , even though meta-analyses using the same functional form suggested that such a representation would lead to consistent emergent F_{CH_4} – T_{air} dependencies among aquatic, wetland, and rice paddy ecosystems³⁰. Considering intra-seasonal variability in emergent F_{CH_4} – T_{air} dependence leads to higher and lower apparent activation energies for F_{CH_4} during earlier and later parts of the frost-free season, respectively (Supplemental Fig. 15a). Our findings indicate that the F_{CH_4} temperature sensitivity is an emergent property that varies substantially with space and time and thus cannot be sufficiently generalized for formulating mechanistic CH₄ models, regardless of its functional form.

Factors other than temperature modulate CH₄ emissions. We applied two approaches to evaluate factors regulating the emergent F_{CH_4} – T_{air} dependence and examine the degree of complexity needed in F_{CH_4} parameterizations in biogeochemical models. In the first approach, we examined the effects of T_{air} , ecosystem-type variability (i.e., differences between ecosystem types), inter-site variability, inter-annual variability, and intra-seasonal variability

on F_{CH_4} predictions. Specifically, F_{CH_4} estimates obtained from six sets of regression models selectively representing the above-mentioned variability (Methods; Supplemental Table 2) were evaluated to investigate how spatial and temporal complexity influences model performance. In the second approach, we trained a random-forest model (Methods) with the FLUXNET-CH₄ database to identify factors controlling the hysteresis parameter a_{hys} (Methods) that quantifies the functional relationship between F_{CH_4} and T_{air} . To assess whether an observationally inferred model can be constructed for F_{CH_4} estimates, we evaluated the predictive power of a hybrid model that uses the random-forest predicted a_{hys} to describe the emergent F_{CH_4} – T_{air} dependence (Methods; Eq. 1) in each part of the frost-free season.

The seven T_{air} dependence models (six regression and one hybrid) can be broadly categorized into three tiers based on the absolute bias relative to the measured F_{CH_4} : (1) employing a universal emergent F_{CH_4} – T_{air} dependence inferred from measurements across the globe without representing spatial and temporal variability (76.2% biased); (2) including ecosystem-type variability (i.e., the emergent F_{CH_4} – T_{air} dependence is inferred from measurements collected at the same ecosystem type, so sites within an ecosystem type are uniformly represented; 63.5–63.9% biased); and (3) including ecosystem-site variability (i.e., the emergent F_{CH_4} – T_{air} dependence is inferred from measurements collected at each site; 38.1–45.9% biased) (Fig. 4). Our results suggest that representing ecosystem-type variability does not necessarily improve F_{CH_4} estimates, because the absolute bias of modeled F_{CH_4} is comparable with that estimated by using a universal emergent F_{CH_4} – T_{air} dependence, except for bog, peat plateau, and wet tundra sites (Fig. 4a). For each ecosystem type, the absolute bias of modeled F_{CH_4} is reduced when ecosystem-site variability is represented, demonstrating the need to recognize inter-annual and inter-site variability (e.g., Fig. 3). For each T_{air} dependence model, the absolute bias of modeled F_{CH_4} is generally higher in rice paddies and salt marshes than in other ecosystem types, suggesting that F_{CH_4} in these systems are sensitive to factors other than T_{air} . For example, timing of irrigation, drainage, planting, and harvesting can all affect F_{CH_4} dynamics in rice paddies⁴⁷.

Results derived from our random-forest model confirm the importance of ecosystem-site variability in regulating a_{hys} and thereby F_{CH_4} predicted by the hybrid model in each part of the frost-free season (Supplemental Fig. 16). Our random-forest predictor importance analysis indicates that site-year specific F_{CH_4} and T_{air} values are more important for a_{hys} estimates than other predictors such as latitude, GPP, and ecosystem type. The weak relationships found between seasonal F_{CH_4} hysteresis and latitude (Supplemental Fig. 14h) and GPP (Supplemental Fig. 14d) are consistent with the relatively low predictor importance for a_{hys} found in our random-forest model. Collectively, our results demonstrate the importance of recognizing inter-site, inter-annual, and intra-seasonal variability for the interpretation of emergent F_{CH_4} – T_{air} dependence inferred from measurements across distinct site-years.

When using a universal emergent F_{CH_4} – T_{air} dependence that only represents a generic T_{air} sensitivity of F_{CH_4} (i.e., the top row in Fig. 4a), the resulting F_{CH_4} predictions substantially underestimate the range of F_{CH_4} measured across wetland and rice paddy sites (Fig. 5a). This generic T_{air} sensitivity of F_{CH_4} flattens the high temporal and spatial variability^{10,11} that strongly controls the timing and magnitude of F_{CH_4} , reinforcing the need

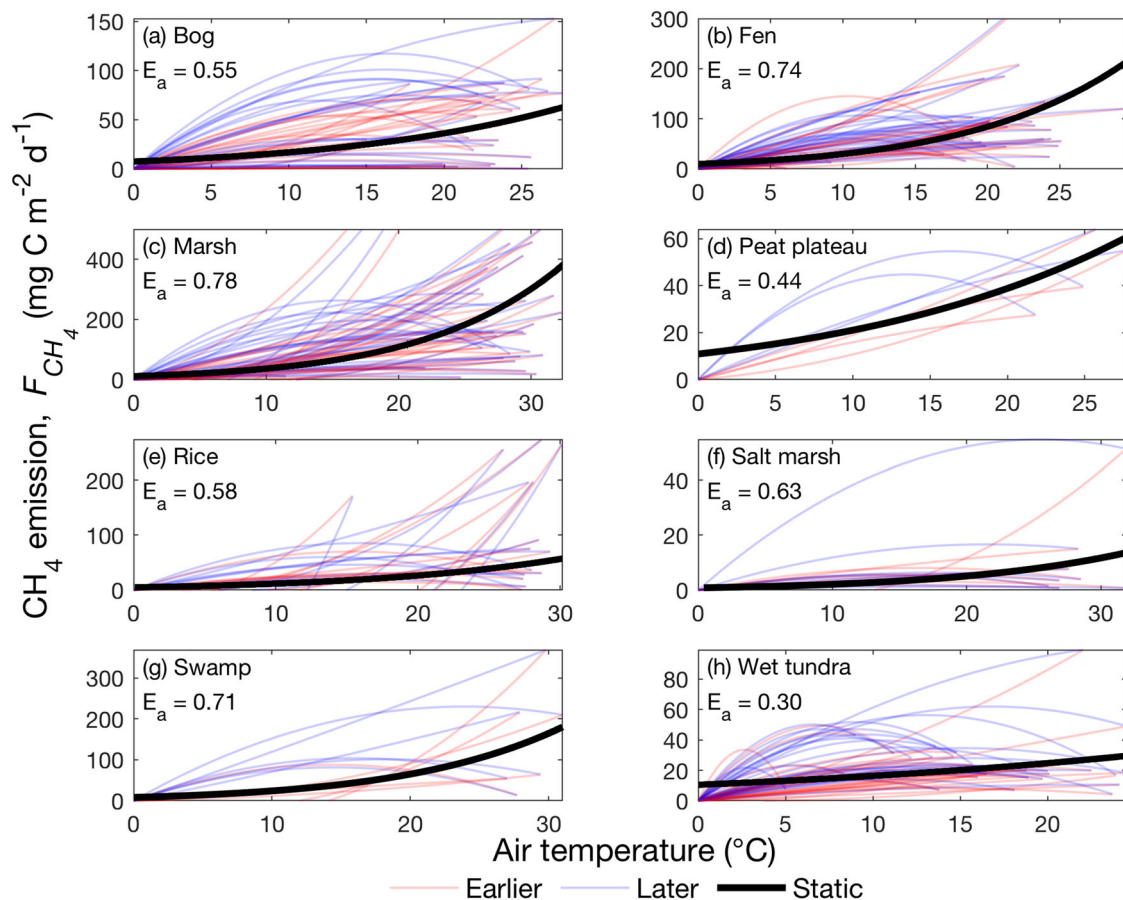


Fig. 3 Large differences in intra-seasonal, inter-annual, and inter-site F_{CH_4} emergent temperature dependencies are found for all examined ecosystem types. Thin lines represent the site- and time-specific emergent dependencies of CH_4 emissions on air temperature inferred from daily measurements collected at bog (a), fen (b), marsh (c), peat plateau (d), rice paddy (e), salt marsh (f), swamp (g), and wet tundra (h) sites. Thick black lines represent ecosystem-type specific emergent dependencies of CH_4 emission on air temperature inferred from the Boltzmann-Arrhenius function that do not recognize spatial heterogeneity and temporal variability. The results inferred from earlier and later parts of the frost-free season, and full frost-free season are colored in red, blue, and black, respectively.

to parameterize factors other than T_{air} in CH_4 models. Including factors other than a generic T_{air} sensitivity of F_{CH_4} (i.e., the bottom row in Fig. 4a) improves F_{CH_4} predictions (Fig. 5b, c), which suggests that F_{CH_4} and emergent $F_{CH_4}-T_{air}$ dependence strongly depend on site- and time-specific environmental conditions. Therefore, models should mechanistically represent CH_4 biogeochemistry, because site- and time-specific emergent $F_{CH_4}-T_{air}$ dependence cannot be accurately parameterized everywhere and all the time. Although many CH_4 models parameterize methanogenesis, methanotrophy, and CH_4 transport for F_{CH_4} modeling¹³, only three of 40 recently reviewed CH_4 models mechanistically represent CH_4 biogeochemistry based on explicit microbial dynamics²⁹. Consequently, implementing process-based representations of CH_4 biogeochemistry in CH_4 models is necessary to improve F_{CH_4} predictions across ecosystem and global scales. Such efforts are imperative because the F_{CH_4} prediction error can increase substantially with increased F_{CH_4} , especially for the relatively simple parameterization that only represents a generic T_{air} sensitivity of F_{CH_4} (Fig. 5c).

Limitations and implications. Additional measurements and analysis of factors controlling methanogenesis, methanotrophy, and CH_4 transport will be needed to investigate the cause of the predominantly positive seasonal F_{CH_4} hysteresis we observed

across wetland and rice paddy sites. When anoxic conditions are prevalent and T_{soil} is the most important driver regulating F_{CH_4} ^{26,27} (e.g., Supplemental Fig. 4), the observed positive seasonal F_{CH_4} hysteresis is consistent with the higher F_{CH_4} driven by higher substrate availability later in the frost-free season²⁵. We identified some environmental drivers affecting the emergent $F_{CH_4}-T_{air}$ dependence at sites where the necessary measurements were available: (1) When WTD drops below the critical zone of CH_4 production later in the frost-free season³¹, the reduced F_{CH_4} may drive negative seasonal F_{CH_4} hysteresis in a given site-year (e.g., the Kopuatai bog in New Zealand (NZ-Kop), Supplemental Fig. 3). (2) F_{CH_4} may become more sensitive to T_{air} changes under higher salinity⁴⁸, and our results indicate that seasonal F_{CH_4} hysteresis shifts from positive to negative with increased salinity (e.g., the Sacramento-San Joaquin Delta of California in USA (US-Myb), Supplemental Fig. 17).

As for the emergent $F_{CH_4}-T_{soil}$ dependence, our results suggest that the functional relationship between F_{CH_4} and T_{soil} may vary non-monotonically along the soil profile. For example, the positive seasonal F_{CH_4} hysteresis inferred from T_{soil} measured at 16 cm depth is stronger than those at 8 and 32 cm depths at US-Myb (Supplemental Fig. 18). Such a non-monotonic relationship indicates that the magnitude of seasonal F_{CH_4} hysteresis is not

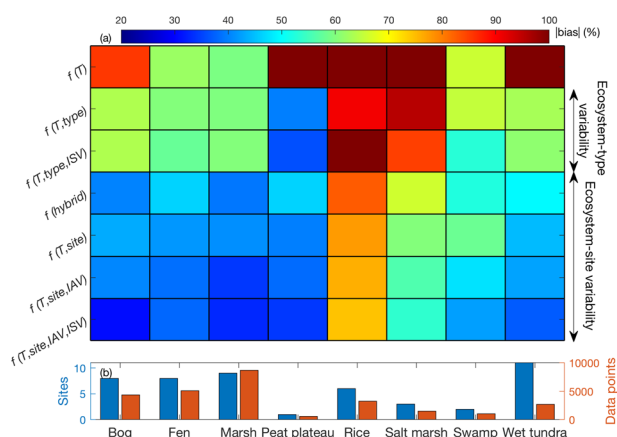


Fig. 4 The accuracy of CH₄ emission estimates improves with better representation of the large wetland-site variability caused by varying environmental conditions. The absolute bias relative to measured CH₄ emissions estimated by each model class for each ecosystem type (a). Blue and red bars denote the number of sites and quality-controlled daily data points within each ecosystem type, respectively (b). The abbreviations used in each model group represent air temperature (T), ecosystem-type variability (type), intra-seasonal variability (ISV), hybrid model based on random-forest estimated hysteresis parameter (hybrid), inter-site variability (site), and inter-annual variability (IAV).

simply caused by time lags between T_{soil} and T_{air} , suggesting that factors other than temperature can strongly control F_{CH_4} . T_{soil} measured at depths where methanogenesis is occurring will be needed to rigorously examine the emergent dependence of F_{CH_4} on T_{soil} across the globe, but such depth-dependent measurements are not yet available among sites in the FLUXNET-CH₄ database. To improve understanding of mechanisms leading to seasonal F_{CH_4} hysteresis, we urge further long-term measurements on factors modulating CH₄ biogeochemistry (e.g., WTD, T_{soil} , microbial activity, and substrate availability), especially in the tropics and the Southern Hemisphere, both of which are sparsely represented in the FLUXNET-CH₄ database. Although seasonal F_{CH_4} hysteresis occurs across seasonal climate and latitudinal gradients (Supplemental Fig. 14), better-representing ecosystems south of 30°N could affect the partitioning of negative and positive seasonal F_{CH_4} hysteresis inferred from existing measurements. While our synthesis in tropical and subtropical regions shows intra-seasonal changes in emergent F_{CH_4} – T_{air} dependence (Supplemental Fig. 19), future studies are needed to examine seasonal F_{CH_4} hysteresis in wetlands south of 30°N (that account for about 75% of global wetland F_{CH_4}).

The observed seasonal F_{CH_4} hysteresis provides a benchmark to evaluate modeled F_{CH_4} functional responses and should inform and motivate CH₄ model development and refinement. Studies have shown that temporal variations in F_{CH_4} are strongly modulated by substrate and microbial dynamics^{33,49,50}, which may explain the substantial seasonal F_{CH_4} hysteresis identified in our wetland and rice paddy sites. For example, a model that explicitly represents substrate and microbial dynamics reproduced the observed hysteretic F_{CH_4} to temperature relationships in several wetlands with different vegetation and hydrological conditions³³. Such dynamics could be parameterized in the terrestrial components of Earth system models⁴⁹. Our synthesis thus provides observational evidence for incorporating substrate and microbial dynamics into next generation CH₄ models.

Using the largest available database of ecosystem-scale CH₄ emissions measured by eddy covariance flux towers, we show that

the apparent relationships between CH₄ emissions and air and soil temperatures are hysteretic and vary strongly with sampling location and measurement period. Approximately 77% of site-years recorded in the wetland and rice paddy subset of the FLUXNET-CH₄ database⁴⁰ show that CH₄ emissions become higher later in the frost-free season at the same air temperature. This predominantly positive seasonal CH₄ emission hysteresis may be driven by substrate-mediated higher CH₄ production²⁵ later in the frost-free season³³. Changes in environmental conditions also modulate seasonal CH₄ emission hysteresis and thus ecosystem-scale CH₄ emissions.

Our results demonstrate that the relationship between CH₄ emissions and temperature is an emergent property that varies substantially across space and time. A direct integration of measurements across the globe (e.g., inferring a generic temperature sensitivity of CH₄ emissions) may not improve CH₄ model parameterization because such an approach oversimplifies factors controlling CH₄ emissions. Therefore, meta-analyses of CH₄ biogeochemistry should recognize the large intra-seasonal, inter-annual, and inter-site variability of biotic and abiotic conditions that regulate ecosystem-scale CH₄ emissions. Collectively, our analyses highlight the importance of observing and modeling spatial heterogeneity and temporal variability for the modeling of CH₄ biogeochemistry. Since most existing CH₄ models are developed using empirically based CH₄ production or emission temperature dependencies²⁹, our study motivates models to mechanistically represent methanogenesis, methanotrophy, and CH₄ transport to refine estimates of global CH₄ emissions and climate feedbacks⁵¹.

Methods

FLUXNET-CH₄ database. The FLUXNET-CH₄ initiative is led by the Global Carbon Project (<https://www.globalcarbonproject.org>) in coordination with regional flux networks (in particular AmeriFlux and the European Fluxes Database) to compile a global CH₄ flux database of eddy covariance and supporting measurements encompassing freshwater, coastal, natural and managed wetlands, and uplands⁴⁰. Database descriptions, including existing sites, data standardization, gap-filling, and partitioning, have been detailed previously in Knox et al.⁴⁰. We used daily mean temperature (air and soil), gross primary productivity as partitioned from net CO₂ exchange measurements, precipitation, WTD, wind speed, atmospheric pressure, and CH₄ emissions compiled at the 48 wetland and rice paddy sites (Supplemental Table 1) currently recorded in the FLUXNET-CH₄ database. Soil temperature is often measured at different depths among different sites, and only about half of the wetland sites report WTD in the current FLUXNET-CH₄ database⁴⁰. We analyzed the soil temperature reported at the shallowest and deepest measured soil layers at each site to investigate their effects on regulating CH₄ emissions. The wetland and rice paddy data (207 site-years with 62,384 site-days as of this publication) were categorized into eight CH₄ emitting ecosystem types: bog, fen, marsh, peat plateau, rice paddy, salt marsh, swamp, and wet tundra, based on previous classification^{52,53}. While gap-filled data are examined, they are not included in our discussion to eliminate potential biases caused by the gap-filling procedure⁴⁵.

Frost-free season. We define the frost-free season as the period when the observed temperature (air or soil) is >0°C to investigate the emergent temperature responses to CH₄ emissions (F_{CH_4}) during the biologically active season across distinct climatic zones. Other data sampling thresholds, such as above-zero GPP and above 5% of annual GPP maximum, were examined, and positive seasonal F_{CH_4} hysteresis is identified in 68–81% of site-years (Supplemental Figs. 20, 21), consistent with those inferred from frost-free season. We chose to present the frost-free season results because substantial GPP (e.g., above 5% of annual GPP maximum) is detected when air temperature is well below 0°C (Supplemental Fig. 22) that may complicate our discussion of varying F_{CH_4} led by temperature changes.

Emergent temperature dependence calculation and the hysteresis parameter

a_{hys} . Emergent dependence of CH₄ emission (F_{CH_4}) on temperature (air or soil) is determined by fitting frost-free-season daily measurements of F_{CH_4} and air and soil temperatures with a quadratic equation (Eq. 1), the Boltzmann–Arrhenius equation (Eq. 2), and first, second, third, and fifth order polynomials. Daily F_{CH_4} estimates made by site- and time-specific emergent F_{CH_4} temperature (air or soil) dependence models based on the above-mentioned functional forms show comparable root mean square errors (Supplemental Fig. 2). Results inferred from the

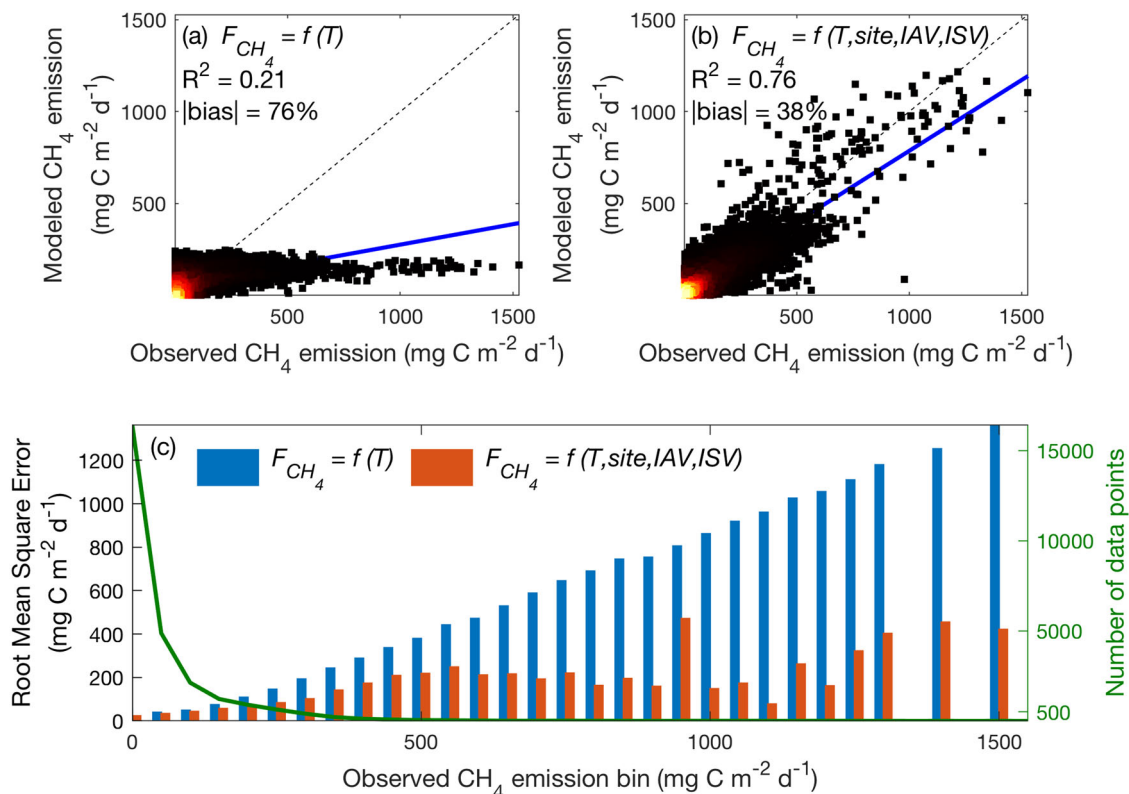


Fig. 5 CH₄ emission prediction error increases substantially as measured CH₄ emission increases. The performance of CH₄ emissions modeled by the regression models that only include a universal emergent CH₄ emission temperature dependence (a), and those that include site- and time-specific conditions (b). The root mean square errors associated with the regression models used in (a) and (b) (bars, left axis) and number of data points (green line, right axis) for measured CH₄ emission bins (c). Two of the 27,130 daily observations have CH₄ emission above 1600 mg C m⁻² d⁻¹, which are not shown for the ease of representation. Lighter colors in the density scatter plot represent denser data points. Solid blue and dashed black lines represent the linear best-fit and one-to-one lines, respectively. The abbreviations used in each model group represent air temperature (T), intra-seasonal variability (ISV), inter-site variability (site), and inter-annual variability (IAV).

quadratic equation (Eq. 1) are selected because (1) its functional form is mathematically consistent with the second-order polynomial equation of temperature for methanogenesis inferred from the MacroMolecular Rate Theory^{54,55}; and (2) it can prescribe seasonal F_{CH_4} hysteresis with a single site- and time- specific parameter (a_{hys} , defined below).

The fits based on the quadratic equation were forced to pass through the origin (assuming zero F_{CH_4} at 0 °C, discussed below) and F_{CH_4} measured at maximum seasonal temperature in each site-year using the Matlab (MathWorks Inc., 2019, version 9.7.0) polyfix function (downloaded from <https://www.mathworks.com/matlabcentral/fileexchange/54207-polyfix-x-y-n-xfix-yfix-xder-dydx>). The resulting emergent dependence of F_{CH_4} on temperature at any given time period can thus be represented as:

$$F_{CH_4}(T) = a_{hys} \bullet T^2 + \left(\frac{F_{CH_4, T_{max}}}{T_{max}} - a_{hys} \bullet T_{max} \right) \bullet T \quad (1)$$

The symbols used in Eq. 1 denote CH₄ emission ($F_{CH_4}(T)$, mg C m⁻² d⁻¹), hysteresis parameter (a_{hys} , mg C m⁻² d⁻¹ °C⁻²), daily mean temperature (T , °C; air or soil), maximum seasonal temperature (T_{max} , °C), and CH₄ emission measured at maximum seasonal temperature ($F_{CH_4, T_{max}}$, mg C m⁻² d⁻¹). Therefore, the functional relationship between and temperature, described by a quadratic equation (Eq. 1), is only determined by the value of hysteresis parameter (a_{hys}) and site-year variables ($F_{CH_4, T_{max}}$ and T_{max}).

The two constraints (passing through the origin and F_{CH_4} measured at maximum seasonal temperature) imposed in Eq. 1 are intended to force the two (earlier and later part of the frost-free season) emergent F_{CH_4} temperature (air or soil) dependencies to form a closed apparent hysteresis loop for each frost-free season. By doing so, seasonal F_{CH_4} hysteresis can be quantified as the normalized area enclosed by the two fits, and intra-seasonal changes can be consistently compared among site-years across distinct climate zones. Ignoring F_{CH_4} around 0 °C has small effects on the magnitude and distribution of seasonal F_{CH_4} hysteresis inferred from the current FLUXNET-CH₄ database, although substantial F_{CH_4} may continue when air temperature is around or below 0 °C^{32,42}. To quantify the effect of ignoring F_{CH_4} around 0 °C, we replaced the constraint of zero F_{CH_4} at 0 °C by the

mean F_{CH_4} measured between -0.5 and 0.5 °C at 0 °C for each site-year, and found that the resulting patterns of seasonal F_{CH_4} hysteresis (Supplemental Fig. 23) are consistent with those assuming zero F_{CH_4} at 0 °C (Fig. 2).

Seasonal CH₄ emission hysteresis. We apply a quadratic equation (Eq. 1) to calculate the emergent dependence of CH₄ emission (F_{CH_4}) on temperature at the earlier ($F_{CH_4, earlier}(T)$) and later ($F_{CH_4, later}(T)$) part of the frost-free season separated by maximum seasonal temperature (T_{max}). Two metrics are used to quantify the observed seasonal F_{CH_4} hysteresis: (1) Normalized area of seasonal F_{CH_4} hysteresis (H_A), defined as the area enclosed by emergent dependencies of F_{CH_4} on temperature inferred from earlier and later parts of the frost-free season (i.e.,

$$H_A = \frac{\int_0^{T_{max}} (F_{CH_4, later}(T) - F_{CH_4, earlier}(T)) dT}{\max(\text{abs}(F_{CH_4, earlier}(T), F_{CH_4, later}(T))) \cdot T_{max}}; \text{ and (2) mean seasonal } F_{CH_4} \text{ hysteresis } (H_\mu),$$

defined as the difference between mean daily F_{CH_4} inferred from measurements taken between later and earlier parts of the frost-free season. In each site-year, positive seasonal F_{CH_4} hysteresis occurs when higher F_{CH_4} are measured later in the frost-free season at a given air or soil temperature. Hysteretic patterns are similar when using either air temperatures (Fig. 1) or soil temperatures (Supplemental Fig. 4), and with either gap-filled (Supplemental Fig. 5) or non-gap-filled (Fig. 1) F_{CH_4} ⁴⁵. Results derived from air temperature (Fig. 2), soil temperature measured at the shallowest soil layer (Supplemental Fig. 12), and soil temperature measured at the deepest soil layer (Supplemental Fig. 13) all indicate predominantly positive seasonal F_{CH_4} hysteresis across the wetland and rice paddy sites. We chose to present results derived from air temperature for its longer and more continuous record in the wetland and rice paddy subset of FLUXNET-CH₄ database, although soil temperature has been shown to be a better predictor for F_{CH_4} ^{33,42}. Specifically, there are 207, 112, and 97 site-years of measurements of air temperature, soil temperature measured at the shallowest soil layer (0–18.3 cm), and soil temperature measured at the shallowest soil layer (32–50 cm), respectively.

Temperature dependence model groups. The measurements extracted from the FLUXNET-CH₄ database were analyzed by seven air temperature (T_{air})

dependence model groups (six regression models and a hybrid model) to evaluate factors modulating CH₄ emission predictions. We design the six regression models to selectively represent the effects of ecosystem-site variability and ecosystem-type variability on CH₄ emission prediction by labeling data points into different groups. The relationship between CH₄ emission and T_{air} is analyzed at each part of the frost-free season, each site-year, each site, and each ecosystem type to quantify intra-seasonal, inter-annual, inter-site, and ecosystem-type variability, respectively (Supplemental Table 2). For the hybrid model, we use the hysteresis parameter predicted by our random-forest model to inform the quadratic equation (Eq. 1) for CH₄ emission estimates. The performance of each T_{air} dependence model group was evaluated to determine the most important model components required for accurate CH₄ emission estimates.

Random-forest model selection. We used random-forest model selection to identify the most important predictors of the hysteresis parameter a_{hys} (Eq. 1) that determines the functional form of emergent CH₄ emission air temperature dependence and thereby wetland CH₄ emissions (F_{CH_4}). Instead of F_{CH_4} , the hysteresis parameter a_{hys} was analyzed, so the results can provide useful information on the source of observed F_{CH_4} hysteresis with an understandable functional form (Eq. 1). Moreover, the most important predictors identified by the machine-learning approach can be compared with the results derived from the other approach using a range of temperature dependence model groups (Supplemental Table 2).

Ten potential predictors were selected for their relatively high predictor importance to a_{hys} : seasonal branch (i.e., earlier or later part in the frost-free season), GPP cumulated in a seasonal branch, precipitation cumulated in a seasonal branch, maximum seasonal temperature, mean temperature in a seasonal branch, ecosystem type, latitude, site, site-year, and F_{CH_4} measured at maximum seasonal temperature. Other potential predictors, including observational year, mean WTD in a seasonal branch, mean wind speed in a seasonal branch, and mean atmospheric pressure in a seasonal branch were examined and showed limited predictive power on a_{hys} . Four potential predictors (seasonal branch, ecosystem type, site, and site-year) were labeled as categorical data and the rest were labeled as numerical data in our random-forest model. The random-forest model selection was performed by the Statistics and Machine-Learning Toolbox in Matlab (MathWorks Inc., 2019, version 9.7.0).

Apparent activation energy for CH₄ emissions. We quantify the apparent activation energy for CH₄ emissions by fitting frost-free-season daily measurements of CH₄ emission and air temperature with the Boltzmann–Arrhenius equation of the form:

$$\ln F_{\text{CH}_4}(T) = \bar{E}_a \bullet \left(\frac{-1}{kT} \right) + \varepsilon \quad (2)$$

where $F_{\text{CH}_4}(T)$ is the rate of CH₄ emission at absolute air temperature T . \bar{E}_a (in eV) and ε correspond to the fitted apparent activation energy (slope) and base reaction rate (intercept), respectively. k is the Boltzmann constant (8.62×10^{-5} eV K⁻¹). When the large inter-site, inter-annual, and intra-seasonal variability is muted, the apparent activation energy for CH₄ emission inferred from each ecosystem type is within the range reported in recent meta-analyses³⁰.

Data availability

This work used publicly available FLUXNET-CH₄ Dataset acquired and shared by the FLUXNET community. All related data is publicly available for download at <https://fluxnet.org/>.

Code availability

Code used in the analysis presented in this study is available online, and can be accessed at https://github.com/ckychang/FCH4_hysteresis⁵⁶.

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References

- Bastviken, D., Tranvik, L. J., Downing, J. A., Crill, P. M. & Enrich-Prast, A. Freshwater methane emissions offset the continental carbon sink. *Science* **331**, 50 (2011).
- Myhre, G. et al. Anthropogenic and Natural Radiative Forcing. in *Climate Change 2013 - The Physical Science Basis* (ed. Intergovernmental Panel on Climate Change) 23, 659–740 (Cambridge University Press, 2013).
- Stocker, B. D. et al. Multiple greenhouse-gas feedbacks from the land biosphere under future climate change scenarios. *Nat. Clim. Change* **3**, 666–672 (2013).
- Saunois, M. et al. The global methane budget 2000–2012. *Earth Syst. Sci. Data* **8**, 697–751 (2016).
- Kirschke, S. et al. Three decades of global methane sources and sinks. *Nat. Geosci.* **6**, 813–823 (2013).
- Saunois, M. et al. The Global Methane Budget 2000–2017. *Earth Syst. Sci. Data* **12**, 1561–1623 (2020).
- Saunois, M. et al. Variability and quasi-decadal changes in the methane budget over the period 2000–2012. *Atmos. Chem. Phys.* **17**, 11135–11161 (2017).
- Saunois, M., Jackson, R. B., Bousquet, P., Poulter, B. & Canadell, J. G. The growing role of methane in anthropogenic climate change. *Environ. Res. Lett.* **11**, 120207 (2016).
- van Vuuren, D. P. et al. The representative concentration pathways: an overview. *Clim. Change* **109**, 5–31 (2011).
- Morin, T. H. et al. Combining eddy-covariance and chamber measurements to determine the methane budget from a small, heterogeneous urban floodplain wetland park. *Agric. Meteorol.* **237–238**, 160–170 (2017).
- Hemes, K. S., Chamberlain, S. D., Eichmann, E., Knox, S. H. & Baldocchi, D. D. A biogeochemical compromise: the high methane cost of sequestering carbon in restored wetlands. *Geophys. Res. Lett.* **45**, 6081–6091 (2018).
- Barba, J. et al. Methane emissions from tree stems: a new frontier in the global carbon cycle. *N. Phytol.* **222**, 18–28 (2019).
- Wania, R. et al. Present state of global wetland extent and wetland methane modelling: methodology of a model inter-comparison project (WETCHIMP). *Geosci. Model Dev.* **6**, 617–641 (2013).
- Melton, J. R. et al. Present state of global wetland extent and wetland methane modelling: conclusions from a model inter-comparison project (WETCHIMP). *Biogeosciences* **10**, 753–788 (2013).
- Bridgman, S. D., Cadillo-Quiroz, H., Keller, J. K. & Zhuang, Q. Methane emissions from wetlands: Biogeochemical, microbial, and modeling perspectives from local to global scales. *Glob. Change. Biol.* **19**, 1325–1346 (2013).
- Hatala, J. A., Detto, M. & Baldocchi, D. D. Gross ecosystem photosynthesis causes a diurnal pattern in methane emission from rice. *Geophys. Res. Lett.* **39**, 1–5 (2012).
- Brown, M. G., Humphreys, E. R., Moore, T. R., Roulet, N. T. & Lafleur, P. M. Evidence for a nonmonotonic relationship between ecosystem-scale peatland methane emissions and water table depth. *J. Geophys. Res. Biogeosci.* **119**, 826–835 (2014).
- Pangala, S. R. et al. Large emissions from floodplain trees close the Amazon methane budget. *Nature* **552**, 230–234 (2017).
- Olefeldt, D., Turetsky, M. R., Crill, P. M. & McGuire, A. D. Environmental and physical controls on northern terrestrial methane emissions across permafrost zones. *Glob. Change. Biol.* <https://doi.org/10.1111/gcb.12071> (2013).
- Perryman, C. R. et al. Thaw transitions and redox conditions drive methane oxidation in a permafrost peatland. *J. Geophys. Res. Biogeosci.* **125**, e2019JG005526 (2020).
- Bergman, I., Klarqvist, M., Nilsson, M. Seasonal variation in rates of methane production from peat of various botanical origins: effects of temperature and substrate quality. *FEMS Microbiol. Ecol.* **33**, 181–189 (2000).
- Juottonen, H., Tuittila, E. S., Juutinen, S., Fritze, H. & Yrjälä, K. Seasonality of rDNA- and rRNA-derived archaeal communities and methanogenic potential in a boreal mire. *ISME J.* **2**, 1157–1168 (2008).
- Aben, R. C. H. et al. Cross continental increase in methane ebullition under climate change. *Nat. Commun.* **8**, 1–8 (2017).
- McCalley, C. K. et al. Methane dynamics regulated by microbial community response to permafrost thaw. *Nature* **514**, 478–481 (2014).
- Chang, K.-Y. et al. Methane production pathway regulated proximally by substrate availability and distally by temperature in a high-latitude mire complex. *J. Geophys. Res. Biogeosci.* **124**, 3057–3074 (2019).
- Rinne, J. et al. Temporal variation of ecosystem scale methane emission from a boreal fen in relation to temperature, water table position, and carbon dioxide fluxes. *Glob. Biogeochem. Cycles* **32**, 1087–1106 (2018).
- Ueyama, M., Yazaki, T., Hirano, T., Futakuchi, Y. & Okamura, M. Environmental controls on methane fluxes in a cool temperate bog. *Agric. Meteorol.* **281**, 107852 (2020).
- Helbig, M., Quinton, W. L. & Sontentag, O. Warmer spring conditions increase annual methane emissions from a boreal peat landscape with sporadic permafrost. *Environ. Res. Lett.* <https://doi.org/10.1088/1748-9326/aa8c85> (2017).
- Xu, X. et al. Reviews and syntheses: four decades of modeling methane cycling in terrestrial ecosystems. *Biogeosciences* **13**, 3735–3755 (2016).
- Yvon-Durocher, G. et al. Methane fluxes show consistent temperature dependence across microbial to ecosystem scales. *Nature* **507**, 488–491 (2014).
- Goodrich, J. P., Campbell, D. I., Roulet, N. T., Clearwater, M. J. & Schipper, L. A. Overriding control of methane flux temporal variability by water table dynamics in a Southern Hemisphere, raised bog. *J. Geophys. Res. Biogeosci.* **120**, 819–831 (2015).
- Zona, D. et al. Cold season emissions dominate the Arctic tundra methane budget. *Proc. Natl Acad. Sci.* **113**, 40–45 (2016).

33. Chang, K.-Y., Riley, W. J., Crill, P. M., Grant, R. F. & Saleska, S. R. Hysteretic temperature sensitivity of wetland CH₄ fluxes explained by substrate availability and microbial activity. *Biogeosciences* **17**, 5849–5860 (2020).
34. Tang, J. & Riley, W. J. Weaker soil carbon–climate feedbacks resulting from microbial and abiotic interactions. *Nat. Clim. Change* **5**, 56–60 (2015).
35. Knorr, K. H. & Blodau, C. Impact of experimental drought and rewetting on redox transformations and methanogenesis in mesocosms of a northern fen soil. *Soil Biol. Biochem.* <https://doi.org/10.1016/j.soilbio.2009.02.030> (2009).
36. Roulet, N. T., Ash, R., Quinton, W. & Moore, T. Methane flux from drained northern peatlands: Effect of a persistent water table lowering on flux. *Glob. Biogeochem. Cycles* <https://doi.org/10.1029/93GB01931> (1993).
37. Bansal, S., Tangen, B. & Finocchiaro, R. Diurnal patterns of methane flux from a seasonal wetland: mechanisms and methodology. *Wetlands* **38**, 933–943 (2018).
38. Herbst, M. et al. Climate and site management as driving factors for the atmospheric greenhouse gas exchange of a restored wetland. *Biogeosciences* **10**, 39–52 (2013).
39. Lai, D. Y. F., Moore, T. R. & Roulet, N. T. Spatial and temporal variations of methane flux measured by autochambers in a temperate ombrotrophic peatland. *J. Geophys. Res. Biogeosci.* **119**, 864–880 (2014).
40. Knox, S. H. et al. FLUXNET-CH₄ synthesis activity: objectives, observations, and future directions. *Bull. Am. Meteorol. Soc.* <https://doi.org/10.1175/BAMS-D-18-0268.1> (2019).
41. Collier, N. et al. The International Land Model Benchmarking (ILAMB) system: design, theory, and implementation. *J. Adv. Model. Earth Syst.* **10**, 2731–2754 (2018).
42. Arndt, K. A. et al. Sensitivity of Methane Emissions to Later Soil Freezing in Arctic Tundra Ecosystems. *J. Geophys. Res. Biogeosci.* <https://doi.org/10.1029/2019jg005242> (2019).
43. Oikawa, P. Y. et al. Unifying soil respiration pulses, inhibition, and temperature hysteresis through dynamics of labile soil carbon and O₂. *J. Geophys. Res. Biogeosci.* **119**, 521–536 (2014).
44. Zhang, Q. et al. The hysteresis response of soil CO₂ to $\ln(t-2)$ concentration and soil respiration to soil temperature. *J. Geophys. Res. G Biogeosci.* **120**, 1605–1618 (2015).
45. Knox, S. H. et al. Biophysical controls on interannual variability in ecosystem-scale CO₂ and CH₄ exchange in a California rice paddy. *J. Geophys. Res. Biogeosci.* **121**, 978–1001 (2016).
46. Wohlfahrt, G. & Galvagno, M. Revisiting the choice of the driving temperature for eddy covariance CO₂ flux partitioning. *Agric. Meteorol.* **237–238**, 135–142 (2017).
47. Hwang, Y. et al. Comprehensive assessments of carbon dynamics in an intermittently-irrigated rice paddy. *Agric. For. Meteorol.* <https://doi.org/10.1016/j.agrformet.2020.107933> (2020).
48. Chamberlain, S. D. et al. Effect of drought-induced salinization on wetland methane emissions, gross ecosystem productivity, and their interactions. *Ecosystems* <https://doi.org/10.1007/s10021-019-00430-5> (2019).
49. Chadburn, S. E. et al. Modeled microbial dynamics explain the apparent temperature sensitivity of wetland methane emissions. *Glob. Biogeochem. Cycles* **34**, e2020GB006678 (2020).
50. Mitra, B. et al. Spectral evidence for substrate availability rather than environmental control of methane emissions from a coastal forested wetland. *Agric. Meteorol.* **291**, 108062 (2020).
51. Dean, J. F. et al. Methane feedbacks to the global climate system in a warmer world. *Rev. Geophys.* **56**, 207–250 (2018).
52. Olefeldt, D., Turetsky, M. R., Crill, P. M. & McGuire, A. D. Environmental and physical controls on northern terrestrial methane emissions across permafrost zones. *Glob. Change. Biol.* **19**, 589–603 (2013).
53. Treat, C. C. et al. Tundra landscape heterogeneity, not interannual variability, controls the decadal regional carbon balance in the Western Russian Arctic. *Glob. Change. Biol.* **24**, 5188–5204 (2018).
54. Liang, L. L. et al. Macromolecular rate theory (MMRT) provides a thermodynamics rationale to underpin the convergent temperature response in plant leaf respiration. *Glob. Change. Biol.* **24**, 1538–1547 (2018).
55. Schipper, L. A., Hobbs, J. K., Rutledge, S. & Arcus, V. L. Thermodynamic theory explains the temperature optima of soil microbial processes and high Q₁₀ values at low temperatures. *Glob. Change. Biol.* **20**, 3578–3586 (2014).
56. Chang, K.-Y. Substantial hysteresis in emergent temperature sensitivity of global wetland CH₄ emissions. GitHub Repository. <https://doi.org/10.5281/zenodo.4571561> (2021).

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Author contributions

All authors contributed to this work. K.Y.C. and W.J.R. designed the analysis, K.Y.C. performed the analysis, and K.Y.C. and W.J.R. analyzed the results and wrote the paper, with inputs from all authors. M.A., D.B., G.B., D.I.C., A.C., A.R.D., E.E., T.F., M.G., M.H., K.S.H., T.H., H.I., M.K., K.W.K., A.L., I.M., B.M., A.M., M.B.N., A.N., W.C.O., M.P., M.L.R., J.R., B.R.K.R., Y.R., T.S., K.V.R.S., H.P.S., N.S., O.S., A.C.I.T., E.S.T., M.U., R.V., T.V., L.W.M., and D.Z. collected the observations. S.H.K., R.B.J., G.M., B.P., S.B., H.C., K.B.D., T.K., D.P., M.T., C.T., and Z.Z. build and maintain the FLUXNET-CH₄ database.

Competing interests

The authors declare no competing interests.

Additional information

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